# PENNSYLVANIA GAME COMMISSION <br> BUREAU OF WILDLIFE MANAGEMENT PROJECT ANNUAL JOB REPORT 

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TITLE: White-tailed Deer Research/Management
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TITLE: Deer Health, Forest Habitat Health, Deer Harvests, and Deer Population Trends by Wildlife Management Unit

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COOPERATING AGENCIES: Pennsylvania Cooperative Fish and Wildlife Research Unit, Pennsylvania Department of Conservation and Natural Resources, Pennsylvania State University, and U.S. Forest Service

WORK LOCATION(S): Statewide
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#### Abstract

We monitored Wildlife Management Unit (WMU) deer health, forest habitat health, and deer population trends using proportion of fawns in the antlerless harvest, advanced tree seedling and sapling regeneration and deer impact from the Pennsylvania Regeneration Study, deer harvest estimates and compositions, and field studies. Proportion of juveniles in the antlerless harvest remained stable in 21 of 22 WMUs and was decreasing in 1 WMU from 2017 to 2022. Forest habitat health was judged to be good in 3 WMUs, fair in 15 WMUs, and poor in 1 WMU. Deer impacts were determined to be acceptable in 17 WMUs and too high in 2 WMUs. Hunters harvested an estimated 422,960 deer (164,190 antlered and 258,770 antlerless), during the 2022-23 deer seasons. Deer populations in 20 WMUs remained stable, while 2 WMUs increased.


## OBJECTIVE

Monitor deer health, forest habitat health, deer harvests, and deer population trends by Wildlife Management Unit (WMU).

## METHODS

## Deer Health

To monitor deer health (i.e., population productivity defined as proportion of fawns in the antlerless harvest), 31 data collection teams examined deer in assigned areas across the state. Each team collected data for 3 days during the first week of the regular firearms season, 2 days during the second week of the season, and 2 days after the close of the season. Data were recorded electronically on Flowfinity software using Apple iPad Minis, and transmitted wirelessly to Flowfinity for analysis. Data collected included age, sex, location of harvest (WMU, county, and township), and hunting license number from ear tags. Deer teams determined deer age as 6 months (fawn), 18 months (yearling), or at least 30 months (adult) using tooth wear and replacement (Severinghaus 1949). Data collection teams also recorded points of antlers and when antlers were physically present, presence or absence of a brow tine on each antler to determine antler characteristics by age class.

We assessed population productivity by monitoring trends in proportion of juveniles in the antlerless harvest (Rosenberry et al. 2011b). We identified proportion of juveniles in the antlerless harvest trends as increasing, decreasing, or stable based on graphical and statistical methods, specifically the Mann-Kendall Test for Trend (Mann 1945, Kendall and Gibbons 1990). We chose this test because it provides a statistical test of trend in data without complex calculations and does not require actual differences between years. Since effective state agency deer programs must consider public involvement and perceptions, it is important that we assess trends with a test that is statistically appropriate, utilizes information available to the public (e.g., a graph of estimates over time), and is relatively easy to explain.

## Forest Habitat Health

We used forest regeneration and deer impacts to assess forest habitat health. Forest regeneration is not just a measure for the benefit of the forest, but also for deer and wildlife. For deer, seedling and sapling trees provide food and cover. As a result, measuring regeneration is an important measure of the sustainability of a forest, and available food and cover that benefit deer and other wildlife.

To obtain data on forest regeneration, advanced tree seedling and sapling regeneration (ATSSR) data are collected as part of a systematic sampling scheme from public and private lands in WMUs from the Pennsylvania Regeneration Study (PRS). This study is being conducted as part of the U.S. Forest Service (USFS) Forest Inventory Analysis in collaboration with Pennsylvania Department of Conservation and Natural Resources (DCNR) and Pennsylvania State University. Subsets of all plots are collected each year, with a complete sampling of plots occurring every 5 7 years. Advanced tree seedling and sapling regeneration from 2 groupings of tree species are available from the PRS. The measure selected for use in deer management is the grouping of dominant canopy species and species capable of achieving high canopy status. "The composition of the ATSSR has a direct impact on the future composition of the forest overstory (Marquis 1994). To cover the range of future forest character and client needs 2 composition groupings are used. The first groups tree species by preference for timber management. The second composition grouping represents the forest's ability to regenerate the existing dominant canopy. Dominant species include those that contribute at least $2 \%$ of the State's total-tree biomass and are able to
grow into the existing canopy; Other High Canopy species include all others that are capable of attaining canopy dominance" (McWilliams et al. 2004).

Based on recommendations from Wildlife Management Institute (Wildlife Management Institute 2010), more plots were included in our analysis of forest regeneration. From 2006 to 2010, only data from plots that were 40 to 75 percent stocked were analyzed. Beginning in 2011, data from all forested plots were analyzed.

We obtained ATSSR data for dominant canopy species and species capable of achieving high canopy status by WMU from the USFS website (USDA 2020) and DCNR. Determination of adequate regeneration was based on levels of deer browse impact observed in the area of each plot. For example, a higher count of seedling and sapling regeneration is required to replace the existing canopy where deer impact is "very high" compared to a lower count of seedling and sapling regeneration where deer impact is "very low." The scaled levels of deer impact indicate deer population size in relation to food availability in a given area. Areas with ample food to support the local deer population will be evident by very low to medium deer impact. Areas lacking food to support the local deer population will be evident by high to very high deer impact. These critical stocking guidelines were derived from extensive literature reviews and decades of research on deerhabitat interactions (Marquis et al. 1992). In 2008 we began using browse impact and associated stocking levels in the habitat health measure. Because of the sampling scheme used in the PRS, it takes 5-7 years to visit all sample plots.

Based on input from cooperating agencies that designed and conduct the PRS and an internal Game Commission review of the forest habitat health measure, we defined forest habitat as "good" if $70 \%$ or more of the sampled plots contained adequate regeneration. If less than $50 \%$ of the plots contained adequate regeneration, forest habitat health was considered "poor." "Fair" falls between levels for "good" and "poor."

Similar to the deer health measure, the forest habitat health measure is based on a sample of plots from across a WMU and we use a statistical test to assess regeneration levels. By using a statistical test to assess differences from predetermined levels (e.g., 70\%), we take into account both the point estimate and associated variation.

When data are collected according to proper sampling design, estimates can be statistically compared to $50 \%$ and $70 \%$ levels using a t-test. The t-test determines whether the estimate is different from the $50 \%$ or $70 \%$ level based on standard statistical procedures. Since reliability of statistical tests is related to sample sizes, forest habitat health determinations are made based on 5year data sets to maximize sample size and reliability of statistical tests.

Decision Rules Used to Determine Forest Habitat Health.--We developed a set of criteria to assign a value of "good," "fair," or "poor" for forest habitat health. A WMU's forest habitat health was considered "good" if the observed percentage of plots with adequate regeneration was greater than, equal to, or not significantly different than 70\%. If a WMU's forest habitat health was not significantly different from $70 \%$ and not significantly different from $50 \%$, then forest habitat health was considered "fair." A WMU’s forest habitat health also was considered "fair" if: 1) the observed percentage of plots with adequate regeneration was equal to $50 \%$; or 2 ) between $50 \%$ and
$70 \%$ and significantly less than $70 \%$; or 3 ) not significantly different than $50 \%$. A WMU's forest habitat health was considered "poor" if the observed percentage of plots with adequate regeneration was significantly less than $50 \%$.

In addition to forest health, we also assessed deer impact on the forest. These data were collected as part of the PRS. Deer impact was assessed on a scale from 1 (very low) to 5 (very high). We identified a score of 3 (moderate) as acceptable deer impact. Similar to the deer and forest health measures, the deer impact measure is based on a sample of plots from across a WMU and we use a statistical test to assess deer impact levels. By using a statistical test to assess differences from predetermined levels (e.g., 3), we take into account both the point estimate and associated variation.

When data are collected according to proper sampling design, estimates can be statistically compared to a score of 3 using a t-test. The t-test determines whether the estimate is different from 3 based on standard statistical procedures. Since reliability of statistical tests is related to sample sizes, deer impact determinations are made based on 5-year data sets to maximize sample size and reliability of statistical tests.

## Deer Harvest Estimates and Composition

To estimate deer harvests and collect data for monitoring deer population trends, 31 data collection teams examined deer in assigned areas across the state. Each team collected data for 3 days during the first week of the regular firearms season, 2 days during the second week of the season, and 2 days after the close of the season. Data were recorded electronically on Flowfinity software using Apple iPad Minis, and transmitted wirelessly to Flowfinity for analysis. Data collected included age, sex, location of harvest (WMU, county, and township), and hunting license number from ear tags. Deer teams determined deer age as 6 months (fawn), 18 months (yearling), or at least 30 months (adult) using tooth wear and replacement (Severinghaus 1949). Data collection teams also recorded points of antlers and when antlers were physically present, presence or absence of a brow tine on each antler to determine antler characteristics by age class.

Data entry for deer harvest report card data was completed by Pennsylvania Game Commission staff. The Pennsylvania Game Commission’s Bureau of Automated Technology Services validated and processed harvest data and ran harvest data analysis programs. For each WMU the analyses included: the number of antlered and antlerless deer checked by aging teams, the number of antlered and antlerless deer checked by deer aging teams and reported by hunters, the total number of antlered and antlerless deer reported by hunters, age and sex composition of the harvest, and reported regular firearms, muzzleloader, and archery harvests.

Deer harvests were estimated using mark-recapture methods. When estimating deer harvests, we used a closed, 2-sample Lincoln-Petersen estimator where deer were considered marked when they were checked in the field by deer aging teams. Recapture occurred when marked deer were reported on report cards, online, or via phone reporting system by hunters.

Because reporting rates in Pennsylvania vary by year, antlered and antlerless deer, and WMU (Rosenberry et al. 2004), deer harvest estimates were calculated for antlered and antlerless deer in each WMU using Chapman's (1951) modified Lincoln-Petersen estimator. This estimator is
recommended (Nichols and Dickman 1996) because it has less bias than the original LincolnPetersen estimator (Chapman 1951).

## Deer Population Trends

We used a modified Sex-Age-Kill (SAK) model to account for Pennsylvania's antler restrictions to monitor deer population trends (i.e., Pennsylvania Sex-Age-Kill [PASAK] model, Norton 2010, Rosenberry et al. 2011a). Modifications involve estimation of 1.5 -year-old and 2.5 -year-old and older male populations. Population trend monitoring relies on research data from Pennsylvania (e.g., Long et al. 2005, Keenan 2010, Norton 2010), harvest estimates, and deer aging data. Population monitoring began with mature males (males 1.5 years of age and older) and progressed to females and fawns. Step-by-step methods and results of the PASAK model were presented to the Board of Commissioners at the January 2011 meeting and posted on the Game Commission's website (Rosenberry et al. 2011a). We also used additional data and further modified the procedure for estimating antlered harvest rates based on age structure of the antlered harvest. This method provided similar population estimates and the benefit of estimates based on annual data rather than multi-year averages used by Norton (2010).

We identified population trends as increasing, decreasing, or stable based on graphical and statistical methods, specifically the Mann-Kendall Test for Trend (Mann 1945, Kendall and Gibbons 1990). We chose this test because it provides a statistical test of trend in data without complex calculations and does not require actual differences between years. Since effective state agency deer programs must consider public involvement and perceptions, it is important that we assess trends with a test that is statistically appropriate, utilizes information available to the public (e.g., a graph of estimates over time), and is relatively easy to explain.

## RESULTS

## Deer Health

Age data from over 16,000 antlerless deer were used to assess proportion of juveniles in the antlerless harvest. Proportion of juveniles in the antlerless harvest ranged from a low of 0.29 in WMUs 2G, 3A, 3C, 4A, and 4D to a high of 0.41 in WMU 1A (Table 1). Twenty-one WMUs showed stable trends, while 1 (WMU 2D) had decreasing trends from 2017 to 2022. An important note is that WMU 2H was dissolved back to its original boundary within WMU 2G in 2022, for a total of 22 WMUs.

## Forest Habitat Health

Wildlife Management Unit level forest habitat health assessments were based on the 5 years of the Pennsylvania Regeneration Study from 2017 to 2022. We identified 3 WMUs (WMUs 2F, 3B, and 5A) with good forest habitat health, and 15 with fair forest habitat health (Table 2). Deer impact was too high in 2 WMUs (WMUs 2C and 3D) and acceptable in 18 WMUs (Table 2). In 3 highly developed WMUs (2B, 5C, and 5D) regeneration and deer impact data were not used or considered in making deer management recommendations because of insufficient sample sizes. Results from this report cannot be compared to some previous years' reports. In reports from 2006 to 2010, only plots with 40 to $75 \%$ stocking levels were analyzed, whereas subsequent reports used all plots.

## Deer Harvest Estimates and Composition

Game Commission personnel checked an average of 286 (range: 60 to 511) antlered deer and 729 (range: 195 to 1,279) antlerless deer per WMU during the 2022 firearms season (Table 3). Based on deer checked and harvest reports by successful hunters, hunters harvested an estimated 422,960 deer in the 2022-23 deer seasons (Table 3). The antlered harvest estimate was 164,190, up $13 \%$ compared to the 2021-22 harvest estimate of 145,320 . The antlerless harvest estimate was 258,770 , up $12 \%$ compared to the harvest estimate of 231,490 in 2021-22.

Antlered harvests were composed of 33\% 1.5-year-old males and 67\% 2.5-year-old and older males (Table 4). Compared to years prior to implementation of antler restrictions during the 2002-03 hunting seasons, the age structure of the antlered harvest has increased, as has the number of 2.5 -year-old and older bucks harvested (Table 4). Antlerless harvest composition has been slowly changing toward more adult females (1.5-year-old and older) since the 1997-98 hunting seasons (Table 5).

## Deer Population Trends

Based on PASAK, deer population trends were stable in 18 WMUs (Table 6). One WMU (2A) had an increasing trend. In WMUs 2B, 5C, and 5D, PASAK cannot be used, but based on antlered harvests and antlerless catch per unit effort estimates, population trends were stable in WMUs 5C and 5D, and increasing in WMU 2B.

## Deer Management Recommendations

Staff evaluate measures of deer health (i.e., proportion of juveniles in the antlerless harvest and population trend), forest habitat health (i.e., percent plots with adequate regeneration), deer impact, and deer-human conflicts from a survey of Pennsylvania citizens (Duda et al. 2019) to develop objectives at the WMU-level (to increase, stabilize, or decrease deer populations). The deer plan objectives include population stabilization in 12 WMUs (1A, 1B, 2A, 2B, 2G, 3A, 3B, 3C, 4C, 5B, 5C, and 5D), and reductions in 10 WMUs (2C, 2D, 2E, 2F, 3D, 4A, 4B, 4D, 4E, and 5A). For WMUs 2A and 2B, where the objective is to stabilize numbers, but there is an increasing population trend, the allocation is set to a level to increase harvest by 1 antlerless deer per square mile above the previous 3 -year average, to stabilize the increasing trend. In WMUs where CWD has been detected in wild deer (WMUs 2C, 2D, 2E, 2F, 4A, 4B, 4D, 4E, and 5A), the objective is to reduce deer populations to reduce CWD transmission and spread, so antlerless allocations have been set to a level to increase antlerless harvest by 1 antlerless deer per square mile above the previous 3-year average. Deer impacts were observed in WMUs 2C and 3D; however, the allocation was already increased in WMU 2C because of CWD. Further, the allocation was already increased in WMU 3D over the past few years because of forest impacts, so the 2023-24 allocation is set to the previous year's level to maintain the previous increase in antlerless harvest level. We continue to recommend consistent regulations that provide more hunting opportunities (e.g., 14 day-concurrent firearms season) and use antlerless allocations to adjust antlerless harvests and population trends. However, it is important to note that as hunter numbers decrease, we anticipate an increased number of WMUs that do not sell out of antlerless licenses or that sell out later in the rounds, but antlerless harvest targets are not being met. A further increase in allocation in these WMUs may not lead to an increase in antlerless harvest. Future efforts to increase antlerless harvest in these WMUs may require additional opportunity (e.g., extended firearms seasons) along with increased antlerless allocations to achieve the needed antlerless harvests. In 2023-24, allocations to address CWD were
based mainly on the WMU level, but we also recommend providing additional DMAP antlerless permits where necessary to enhance surveillance.

## Action by the Board of Commissioners (BOC)

Starting with the 2023-24 antlerless deer license allocations, the Board of Commissioners (BOC) will be discontinuing the practice of annually voting on, potentially changing, and approving the number of antlerless deer licenses to be allocated. Thus, the recommended allocations by staff to meet publicly identified and supported deer plan goals and objectives will move forward as proposed (Table 7). The BOC voted to retain the season-long concurrent firearms season for antlered and antlerless deer season in all WMUs. The BOC voted to keep the opening day of rifle deer season on the Saturday after Thanksgiving, and to allow deer hunting on the Sunday following opening day. The last Sunday of the archery season was again approved as a hunting day. The fall archery season extension was again approved, ending on the Friday before bear season. The limit of DMAP permits per hunter per DMAP area was maintained at 4 (if the landowner provides coupons), and up to two permits in units that the Game Commission designates (offered online with no coupon). The BOC also voted to retain the personal limit of up to 6 unfilled, WMU-specific antlerless licenses at any given time. If one is used, they may purchase another, as long as licenses are available.

## RECOMMENDATIONS

1. Identify and develop additional analyses and measurements to improve the forest habitat health measure's ability to account for factors other than deer that affect forest regeneration and to most directly monitor deer impacts on forest regeneration.
2. Maintain deer aging sampling effort. Current numbers of deer checked in the field provide precise harvest estimates in most WMUs. Harvest estimates are less precise in smaller WMUs where it is more difficult to collect sufficient data.
3. Continue to evaluate validity of assumptions and population monitoring procedures through internal review and analyses and external peer review. Prioritize research needs based on internal and external reviews.
4. Investigate alternatives to the current non-parametric tests to determine trends in current metrics.
5. Continue antler restriction regulations in accordance with goals and objectives of the deer management plan.
6. Continue to allow hunters to purchase and use the entire antlerless allocation.
7. In WMUs containing CWD-positive deer in the free-ranging population, continue to allocate antlerless licenses to reduce the deer population and use DMAP permits to further reduce deer numbers in specific areas where CWD-positive deer have been detected. Chronic wasting disease is rapidly increasing and spreading. Reducing deer populations is the most practical management option at this time.
8. Set antlerless license allocations to achieve deer management goals as defined in the deer management plan. Investigate alternative options when the antlerless allocation is no longer able to meet harvest targets (e.g., not selling out, or an increase in allocation does not lead to an increase in antlerless harvest to target levels).

## LITERATURE CITED

Chapman, D. G. 1951. Some properties of the hypergeometric distribution with applications to zoological censuses. University of California Publications on Statistics 1:131-160.

Duda, M. D., M. Jones, T. Beppler, S. J. Bissell, A. Criscione, P. Doherty, A. Ritchie, C. L. Schilli, T. Winegord, and A. Lanier. 2019. Pennsylvania residents’ opinions on and attitudes toward deer and deer management. Responsive Management National Office, Harrisonburg, Virginia, USA.

Keenan, M. T. 2010. White-tailed deer harvest rate and hunter distribution. Thesis, Pennsylvania State University, University Park, USA.

Kendall, M. G., and J. D. Gibbons. 1990. Rank Correlation Methods. Fifth edition. Edward Arnold, London, United Kingdom.

Long, E. S., D. R. Diefenbach, C. S. Rosenberry, B. D. Wallingford, and M. D. Grund. 2005. Landscape structure influences dispersal distances of a habitat generalist, the white-tailed deer. Journal of Mammalogy 86:623-629.

Mann, H. B. 1945. Non-parametric tests against trend. Econometrica 13:245-259.
Marquis, D. A., R. L. Ernst, and S. L. Stout. 1992. Prescribing silvicultural treatments in hardwood stands of the Alleghenies. Revised editor. U.S. Forest Service General Technical Report NE-96.

Marquis, D. A., editor. 1994. Quantitative silviculture for hardwood forests of the Alleghenies. General Technical Report. NE-183. U.S. Department of Agriculture Forest Service, Northeastern Research Station, Radnor, Pennsylvania, USA.

McWilliams, W. H., C. A. Alerich, D. A. Devlin, A. J. Lister, T. W. Lister, S. L. Sterner, and J. A. Westfall. 2004. Annual inventory report for Pennsylvania's forests: results from the first three years. Resource Bulletin NE-159. USDA Forest Service, Newtown Square, Pennsylvania, USA.

Nichols, J. D. and C. R. Dickman. 1996. Capture-recapture methods. Pages 217-226 in D. E. Wilson, F. R. Cole, J. D. Nichols, R. Rudran, and M. S. Foster, editors. Measuring and monitoring biological diversity: standard methods for mammals. Smithsonian Institute Press, Washington D.C., USA.

Norton, A. S. 2010. An evaluation of the Pennsylvania sex-age-kill model for white-tailed deer. Thesis, Pennsylvania State University, University Park, USA.

Rosenberry, C. S., D. R. Diefenbach, and B. D. Wallingford. 2004. Reporting rate variability and precision of white-tailed deer harvest estimates in Pennsylvania. Journal of Wildlife Management 68:860-869.

Rosenberry, C. S., J. T. Fleegle, and B. D. Wallingford. 2011a. Monitoring deer populations in Pennsylvania. Pennsylvania Game Commission, Harrisburg, USA.

Rosenberry, C. S., A. S. Norton, D. R. Diefenbach, J. T. Fleegle, and B. D. Wallingford. 2011b. White-tailed deer age ratios as herd management and predator impact measures in Pennsylvania. Wildlife Society Bulletin 35:461-468.

Severinghaus, C. W. 1949. Tooth development and wear as criteria of age in white-tailed deer. Journal of Wildlife Management 13:195-216.

United States Department of Agriculture [USDA]. 2020. United State Forest Service. Forest Inventory and Analysis. https://www.fia.fs.fed.us/. Accessed 15 January 2021.

Wildlife Management Institute. 2010. The deer management program of the Pennsylvania Game Commission: a comprehensive review and evaluation. The Wildlife Management Institute, Washington D.C., USA. http://lbfc.legis.state.pa.us/Resources/Documents/Reports/307.pdf. Accessed 22 Oct 2010.

Table 1. Number of antlerless deer examined in 2022, proportion of juveniles in the antlerless 2022 harvest, and trend in the proportion of juveniles in the antlerless harvest by Wildlife Management Unit (WMU) from 2017 to 2022, Pennsylvania.

| WMU | $\boldsymbol{n}$ | Proportion of <br> juveniles in antlerless <br> harvest | Trend |
| :---: | :---: | :---: | :---: |
| 1A | 682 | 0.41 | Stable |
| 1B | 1,205 | 0.34 | Stable |
| 2A | 697 | 0.30 | Stable |
| 2B | 472 | 0.35 | Stable |
| 2C | 923 | 0.36 | Stable |
| 2D | 1,291 | 0.32 | Decreasing |
| 2E | 598 | 0.34 | Stable |
| 2F | 980 | 0.31 | Stable |
| 2G | 577 | 0.29 | Stable |
| 3A | 619 | 0.29 | Stable |
| 3B | 816 | 0.32 | Stable |
| 3C | 896 | 0.29 | Stable |
| 3D | 663 | 0.30 | Stable |
| 4A | 472 | 0.29 | Stable |
| 4B | 597 | 0.33 | Stable |
| 4C | 840 | 0.32 | Stable |
| 4D | 917 | 0.29 | Stable |
| 4E | 1,208 | 0.33 | Stable |
| 5A | 221 | 0.30 | Stable |
| 5B | 1,100 | 0.39 | Stable |
| 5C | 746 | 0.39 | Stable |
| 5D | 290 | 0.36 | Stable |

[^0]Table 2. Number of regeneration plots sampled, percent with adequate regeneration, mean deer impact and qualitative assessments of regeneration and deer impact by Wildlife Management Unit (WMU). Data are based on samples collected from 2017 to 2021, Pennsylvania. Results are based on all forested plots and cannot be compared to results from 2006 to 2010 that only included $40 \%$ to $75 \%$ stocked plots.

| WMU | n | \% plots with adequate regeneration | Forest health assessment | Mean deer impact | Impact assessment |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1A | 16 | 65\% | Fair | 2.9 | Acceptable |
| 1B | 10 | 55\% | Fair | 2.7 | Acceptable |
| 2A | 19 | 34\% | Fair | 3.3 | Acceptable |
| 2B | $\mathrm{n} / \mathrm{a}^{\mathrm{a}}$ | n/aa | $\mathrm{n} / \mathrm{a}^{\text {a }}$ | $\mathrm{n} / \mathrm{a}^{\text {a }}$ | $\mathrm{n} / \mathrm{a}^{\mathrm{a}}$ |
| 2C | 39 | 52\% | Fair | 3.2 | Too high |
| 2D | 26 | 36\% | Poor | 3.2 | Acceptable |
| 2E | 22 | 54\% | Fair | 3.2 | Acceptable |
| 2F | 19 | 69\% | Good | 3.0 | Acceptable |
| $2 \mathrm{G}^{\text {b }}$ | 61 | 52\% | Fair | 2.9 | Acceptable |
| 3A | 12 | 66\% | Fair | 2.8 | Acceptable |
| 3B | 34 | 66\% | Good | 2.8 | Acceptable |
| 3C | 26 | 50\% | Fair | 3.3 | Acceptable |
| 3D | 34 | 53\% | Fair | 3.6 | Too high |
| 4A | 23 | 47\% | Fair | 3.0 | Acceptable |
| 4B | 21 | 52\% | Fair | 3.1 | Acceptable |
| 4C | 21 | 53\% | Fair | 3.2 | Acceptable |
| 4D | 41 | 52\% | Fair | 2.9 | Acceptable |
| 4E | 17 | 49\% | Fair | 3.2 | Acceptable |
| 5A | 4 | 73\% | Good | 3.3 | Acceptable |
| 5B | 9 | 57\% | Fair | 3.1 | Acceptable |
| 5C | $\mathrm{n} / \mathrm{a}^{\mathrm{a}}$ | $\mathrm{n} / \mathrm{a}^{\text {a }}$ | $\mathrm{n} / \mathrm{a}^{\text {a }}$ | $\mathrm{n} / \mathrm{a}^{\text {a }}$ | $\mathrm{n} / \mathrm{a}^{\mathrm{a}}$ |
| 5D | $\mathrm{n} / \mathrm{a}^{\mathrm{a}}$ | $\mathrm{n} / \mathrm{a}^{\text {a }}$ | $\mathrm{n} / \mathrm{a}^{\text {a }}$ | $\mathrm{n} / \mathrm{a}^{\text {a }}$ | $\mathrm{n} / \mathrm{a}^{\text {a }}$ |

${ }^{\text {a }}$ Regeneration data from these highly developed WMUs were not analyzed or considered in making deer management recommendations.
${ }^{\text {b }}$ From 2013-2021, WMU 2G was split into WMUs 2H and 2G. In 2022, WMU 2H was dissolved into WMU 2G.

Table 3. Number of deer checked by Pennsylvania Game Commission personnel, number of deer reported by successful hunters, and estimated harvests for antlered and antlerless deer by Wildlife Management Unit (WMU), Pennsylvania, 2022-23 (excluding DMAP harvests).

|  | Antlered |  |  |  | Antlerless |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| WMU | Deer checked | Reported | Harvest $^{\mathbf{a}}$ |  | Deer checked | Reported | Harvest $^{\text {a }}$ |
| 1A | 196 | 2337 | 9,000 |  | 677 | 4191 | 13,800 |
| 1B | 351 | 2746 | 9,100 |  | 1151 | 4065 | 15,300 |
| 2A | 222 | 2461 | 8,700 |  | 704 | 3341 | 11,000 |
| 2B | 93 | 2104 | 6,600 |  | 485 | 3703 | 15,000 |
| 2C | 355 | 3410 | 10,000 |  | 930 | 5123 | 16,600 |
| 2D | 349 | 4133 | 14,000 |  | 1279 | 7088 | 23,000 |
| 2E | 253 | 2193 | 6,700 |  | 578 | 3581 | 10,600 |
| 2F | 439 | 3280 | 8,800 |  | 841 | 3638 | 11,800 |
| 2G | 379 | 3008 | 8,600 |  | 470 | 2220 | 6,900 |
| 3A | 356 | 1818 | 5,700 |  | 540 | 1816 | 5,600 |
| 3B | 379 | 2408 | 7,300 |  | 770 | 2814 | 8,900 |
| 3C | 511 | 2773 | 8,000 |  | 859 | 3485 | 12,000 |
| 3D | 317 | 2100 | 5,500 |  | 646 | 2739 | 7,400 |
| 4A | 153 | 1508 | 3,800 |  | 486 | 2768 | 11,100 |
| 4B | 186 | 1715 | 4,800 |  | 571 | 2594 | 8,400 |
| 4C | 353 | 2767 | 6,900 |  | 812 | 3052 | 8,200 |
| 4D | 404 | 2736 | 7,900 |  | 847 | 3563 | 12,200 |
| 4E | 312 | 2450 | 8,000 |  | 1123 | 3903 | 12,400 |
| 5A | 60 | 1488 | 3,100 |  | 195 | 2486 | 7,400 |
| 5B | 275 | 3315 | 10,900 |  | 1031 | 5587 | 16,300 |
| 5C | 258 | 2640 | 7,200 |  | 755 | 5444 | 16,700 |
| 5D | 89 | 1290 | 2,500 |  | 285 | 2995 | 6,700 |
| Unk. |  | 371 | 1,090 |  | 499 | 1,470 |  |

[^1]Table 4. Number of yearling (1.5) and adult (2.5-year-old and older) male deer aged, age composition of harvests, and estimated number of 2.5 -year-old and older males harvested in Pennsylvania, 1997-98 to 2022-23. Percentages may not add up to 100 percent due to rounding.

| Year | n | \% 1.5- <br> year-old <br> males | \% 2.5-year- <br> old and <br> older males | Estimate of 2.5-year-old <br> and older males <br> harvested |
| :---: | ---: | :---: | :---: | :---: |
| $1997-98$ | 18,563 | 81 | 19 | 33,600 |
| $1998-99$ | 21,350 | 81 | 19 | 34,500 |
| $1999-00$ | 20,011 | 80 | 20 | 38,900 |
| $2000-01$ | 22,145 | 82 | 18 | 36,600 |
| $2001-02$ | 18,893 | 78 | 22 | 44,700 |
| $2002-03^{\text {a }}$ | 11,694 | 68 | 32 | 52,900 |
| $2003-04$ | 11,367 | 56 | 44 | 62,600 |
| $2004-05$ | 10,559 | 50 | 50 | 62,000 |
| $2005-06$ | 9,062 | 52 | 48 | 57,800 |
| $2006-07$ | 10,819 | 56 | 44 | 59,500 |
| $2007-08$ | 8,014 | 56 | 44 | 48,000 |
| $2008-09$ | 9,357 | 52 | 48 | 59,200 |
| $2009-10$ | 8,443 | 49 | 51 | 55,200 |
| $2010-11$ | 9,032 | 48 | 52 | 64,400 |
| $2011-12^{\text {a }}$ | 10,311 | 50 | 50 | 63,800 |
| $2012-13$ | 10,588 | 48 | 52 | 69,000 |
| $2013-14$ | 9,937 | 47 | 53 | 71,200 |
| $2014-15$ | 9,225 | 43 | 57 | 68,000 |
| $2015-16$ | 9,762 | 41 | 59 | 81,200 |
| $2016-17$ | 9,792 | 44 | 56 | 83,400 |
| $2017-18$ | 11,404 | 43 | 57 | 93,400 |
| $2018-19$ | 9,485 | 36 | 64 | 94,600 |
| $2019-20$ | 8,420 | 34 | 66 | 107,700 |
| $2020-21$ | 7,591 | 36 | 64 | 111,900 |
| $2021-22$ | 6,746 | 38 | 62 | 90,00 |
| $2022-23$ | 6,987 | 33 | 67 | 106,812 |

[^2]Table 5. Number of antlerless deer aged and age composition of harvests in Pennsylvania, 199798 to 2022-23. Percentages may not add up to 100 percent due to rounding.

| Year | $\boldsymbol{n}$ | \% 0.5-year- <br> old males | \% 0.5-year- <br> old females | \% 1.5-year-old and <br> older females |
| :---: | :---: | :---: | :---: | :---: |
| $1997-98$ | 28,743 | 24 | 20 | 56 |
| $1998-99$ | 24,913 | 23 | 20 | 57 |
| $1999-00$ | 18,502 | 24 | 20 | 56 |
| $2000-01$ | 30,460 | 22 | 20 | 58 |
| $2001-02$ | 25,450 | 22 | 18 | 60 |
| $2002-03$ | 30,077 | 22 | 18 | 60 |
| $2003-04$ | 28,236 | 21 | 18 | 61 |
| $2004-05$ | 24,640 | 22 | 18 | 61 |
| $2005-06$ | 19,459 | 23 | 19 | 58 |
| $2006-07$ | 1,074 | 23 | 19 | 58 |
| $2007-08$ | 17,770 | 24 | 20 | 56 |
| $2008-09$ | 17,152 | 22 | 18 | 60 |
| $2009-10$ | 16,519 | 22 | 18 | 60 |
| $2010-11$ | 14,837 | 23 | 18 | 59 |
| $2011-12$ | 16,050 | 21 | 19 | 60 |
| $2012-13$ | 15,563 | 22 | 18 | 61 |
| $2013-14$ | 15,924 | 21 | 18 | 62 |
| $2014-15$ | 14,909 | 20 | 18 | 61 |
| $2015-16$ | 14,551 | 20 | 17 | 63 |
| $2016-17$ | 14,966 | 20 | 16 | 64 |
| $201-18$ | 15,310 | 19 | 17 | 64 |
| $2018-19$ | 15,008 | 17 | 17 | 66 |
| $2019-20$ | 15,104 | 16 | 15 | 69 |
| $2020-21$ | 16,844 | 17 | 15 | 68 |
| $2021-22$ | 15,926 | 16 | 15 | 69 |
| $2022-23$ | 16,810 | 17 | 16 | 67 |

Table 6. Pennsylvania Sex-Age-Kill (PASAK) model estimates of post-hunt deer populations by Wildlife Management Unit (WMU), 2013 to 2023, Pennsylvania.

| WMU | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 | 6-yr Trend |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1A | 48,472 | 55,114 | 49,169 | 62,237 | 65,707 | 53,244 | 46,208 | 51,804 | 99,568 | 57,982 | 73,334 | Stable |
| 1B | 55,713 | 53,799 | 47,438 | 71,669 | 74,053 | 81,376 | 60,756 | 81,659 | 95,277 | 74,887 | 72,506 | Stable |
| 2A | 53,996 | 43,379 | 30,033 | 48,723 | 57,963 | 46,361 | 44,587 | 61,486 | 72,156 | 65,676 | 77,599 | Increasing |
| 2B | a | a | a | a |  | a | , | a | a |  |  | , |
| 2C | 61,386 | 68,683 | 66,027 | 83,350 | 69,034 | 113,659 | 85,400 | 97,246 | 76,365 | 73,906 | 86,600 | Stable |
| 2D | 113,774 | 144,084 | 110,214 | 117,823 | 112,499 | 140,281 | 104,622 | 114,679 | 93,498 | 99,753 | 107,353 | Stable |
| 2E | 44,546 | 45,529 | 50,549 | 43,081 | 43,144 | 56,635 | 46,170 | 62,753 | 52,578 | 54,143 | 56,405 | Stable |
| 2F | 83,063 | 65,614 | 61,020 | 67,152 | 74,387 | 108,575 | 86,836 | 98,104 | 112,840 | 86,470 | 83,968 | Stable |
| $2 \mathrm{G}^{\text {b }}$ | 73,375 | 65,850 | 57,215 | 80,951 | 83,646 | 120,406 | 74,138 | 96,260 | 128,416 | 98,923 | 107,504 | Stable |
| 3A | 41,358 | 45,317 | 36,181 | 49,307 | 49,426 | 55,441 | 39,832 | 54,040 | 71,376 | 55,494 | 59,595 | Stable |
| 3B | 53,709 | 63,803 | 55,249 | 76,808 | 80,598 | 76,249 | 51,976 | 62,489 | 90,795 | 56,589 | 74,283 | Stable |
| 3C | 67,720 | 58,925 | 67,997 | 83,206 | 85,083 | 79,925 | 57,169 | 75,360 | 94,807 | 61,771 | 69,345 | Stable |
| 3D | 29,225 | 25,127 | 33,778 | 28,957 | 33,302 | 30,727 | 33,798 | 48,663 | 45,355 | 32,058 | 52,788 | Stable |
| 4A | 36,579 | 42,196 | 23,772 | 48,538 | 29,746 | 39,238 | 40,344 | 47,047 | 39,911 | 35,442 | 19,763 | Stable |
| 4B | 52,903 | 50,517 | 45,362 | 57,846 | 55,941 | 52,407 | 50,136 | 54,044 | 44,691 | 26,808 | 43,771 | Stable |
| 4C | 45,586 | 49,072 | 50,265 | 55,068 | 55,311 | 61,317 | 55,122 | 55,238 | 77,639 | 52,314 | 64,683 | Stable |
| 4D | 67,011 | 61,428 | 56,905 | 60,398 | 63,984 | 99,997 | 61,441 | 71,983 | 89,963 | 66,855 | 67,514 | Stable |
| 4E | 48,318 | 50,707 | 59,206 | 64,923 | 62,285 | 70,064 | 60,055 | 59,120 | 77,399 | 67,325 | 67,790 | Stable |
| 5A | 28,014 | 29,715 | 25,032 | 20,081 | 28,581 | 33,243 | 25,162 | 49,801 | 28,772 | 20,313 | 21,887 | Stable |
| 5B | 75,260 | 63,591 | 60,538 | 66,282 | 73,573 | 85,790 | 77,485 | 76,623 | 91,713 | 62,401 | 101,325 | Stable |
| 5C | a | a | a | , | , | a | a |  |  | , | , | a |
| 5D | a | a | a | a | a | a | a | a | a | a | a | a |

${ }^{\text {a }}$ PASAK model estimates are not available for these WMUs. See Rosenberry et al. 2011 for further information.
Population trend assessment in these WMUs is based on antlered harvests and antlerless catch per unit effort estimates.
${ }^{\text {b }}$ From 2013-2021, WMU 2G was split into WMUs 2H and 2G. In 2022, WMU 2H was dissolved into WMU 2G.

Table 7. Antlerless license allocations by Wildlife Management Unit (WMU), 2013-14 to 2023-24, Pennsylvania.

| WMU | $\mathbf{2 0 1 3 - 1 4}$ | $\mathbf{2 0 1 4 - 1 5}$ | $\mathbf{2 0 1 5 - 1 6}$ | $\mathbf{2 0 1 6 - 1 7}$ | $\mathbf{2 0 1 7 - 1 8}$ | $\mathbf{2 0 1 8 - 1 9}$ | $\mathbf{2 0 1 9 - 2 0}$ | $\mathbf{2 0 2 0 - 2 1}$ | $\mathbf{2 0 2 1 - 2 2}$ | $\mathbf{2 0 2 2 - 2 3}$ | $\mathbf{2 0 2 3 - 2 4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1A | 49,000 | 47,000 | 46,000 | 46,000 | 52,000 | 48,000 | 49,000 | 49,000 | 40,000 | 43,000 | 46,000 |
| 1B | 31,000 | 30,000 | 29,000 | 29,000 | 35,000 | 37,000 | 35,000 | 41,000 | 32,000 | 34,000 | 37,000 |
| 2A | 49,000 | 46,000 | 43,000 | 43,000 | 50,000 | 49,000 | 46,000 | 46,000 | 39,000 | 39,000 | 46,000 |
| 2B | 62,000 | 60,000 | 61,000 | 61,000 | 60,000 | 58,000 | 54,000 | 49,000 | 49,000 | 49,000 | 53,000 |
| 2C | 43,000 | 38,000 | 31,000 | 31,000 | 31,000 | 44,000 | 52,000 | 58,000 | 67,000 | 67,000 | 88,000 |
| 2D | 61,000 | 61,000 | 55,000 | 55,000 | 55,000 | 63,000 | 66,000 | 60,000 | 74,000 | 74,000 | 86,000 |
| 2E | 22,000 | 21,000 | 21,000 | 21,000 | 22,000 | 27,000 | 32,000 | 39,000 | 42,000 | 42,000 | 52,000 |
| 2F | 29,000 | 27,000 | 22,000 | 22,000 | 24,000 | 23,000 | 31,000 | 36,000 | 32,000 | 37,000 | 49,000 |
| 2G | 34,000 | 27,500 | 28,500 | 27,000 | 32,500 | 36,000 | 32,000 | 34,000 | 32,000 | 31,000 | 35,000 |
| 3A | 23,000 | 18,000 | 19,000 | 15,000 | 20,000 | 22,000 | 20,000 | 21,000 | 19,000 | 19,000 | 21,000 |
| 3B | 39,000 | 33,000 | 28,000 | 28,000 | 30,000 | 29,000 | 38,000 | 33,000 | 30,000 | 33,000 | 32,000 |
| 3C | 35,000 | 32,000 | 36,000 | 36,000 | 42,000 | 38,000 | 46,000 | 49,000 | 33,000 | 37,000 | 40,000 |
| 3D | 32,000 | 25,000 | 25,000 | 25,000 | 25,000 | 25,000 | 25,000 | 36,000 | 36,000 | 41,000 | 41,000 |
| 4A | 28,000 | 28,000 | 30,000 | 30,000 | 30,000 | 38,000 | 41,000 | 49,000 | 50,000 | 50,000 | 61,000 |
| 4B | 24,000 | 26,000 | 26,000 | 26,000 | 26,000 | 26,000 | 32,000 | 33,000 | 34,000 | 34,000 | 46,000 |
| 4C | 27,000 | 25,000 | 25,000 | 25,000 | 29,000 | 30,000 | 36,000 | 32,000 | 29,000 | 31,000 | 32,000 |
| 4D | 35,000 | 33,000 | 33,000 | 34,000 | 34,000 | 34,000 | 46,000 | 45,000 | 55,000 | 55,000 | 77,000 |
| 4E | 26,000 | 21,000 | 25,000 | 25,000 | 27,500 | 32,000 | 34,000 | 37,000 | 42,000 | 42,000 | 54,000 |
| 5A | 19,000 | 19,000 | 19,000 | 19,000 | 22,000 | 23,000 | 22,000 | 26,000 | 31,000 | 31,000 | 40,000 |
| 5B | 50,000 | 49,000 | 50,000 | 50,000 | 57,000 | 58,000 | 67,000 | 60,000 | 60,000 | 60,000 | 60,000 |
| 5C | 103,000 | 95,000 | 70,000 | 70,000 | 70,000 | 70,000 | 70,000 | 70,000 | 70,000 | 70,000 | 70,000 |
| 5D | 18,000 | 18,000 | 24,000 | 30,000 | 30,000 | 28,000 | 29,000 | 29,000 | 29,000 | 29,000 | 29,000 |

[^3]
[^0]:    ${ }^{\mathrm{a}}$ WMU 2G is composed of the former WMUs 2H and 2G. In 2022, WMU 2H was dissolved into WMU 2G.

[^1]:    ${ }^{\text {a }}$ Estimated harvests are rounded to the nearest 100 or 1,000 based on precision of harvest estimate. Unknown WMU harvests are rounded to the nearest 10 due to the small number.
    ${ }^{\text {b }}$ From 2013-2021, WMU 2G was split into WMUs 2H and 2G. In 2022, WMU 2H was dissolved into WMU 2G.

[^2]:    ${ }^{\text {a }}$ Three and 4-point antler restrictions started in 2002-03.
    ${ }^{\mathrm{b}}$ In 2011, the 4-point antler restriction was modified to 3-points not including the brow tine.

[^3]:    ${ }^{\text {a }}$ From 2013-2021, WMU 2G was split into WMUs 2H and 2G. In 2022, WMU 2H was dissolved into WMU 2G.

